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Impact of human intervention and climate change on natural flow regime

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Abstract

According to the ‘natural flow paradigm’, any departure from the natural flow condition will alter the river ecosystem. River flow regimes have been modified by anthropogenic interventions and climate change is further expected to affect the biotic interactions and the distribution of stream biota by altering streamflow. This study aims to evaluate the hydrologic alteration caused by dam construction and climatic changes in a mesoscale river basin, which is prone to both droughts and monsoonal floods. To analyse the natural flow regime, 15 years of observed streamflow (1950-1965) prior to dam construction is used. Future flow regime is simulated by a calibrated hydrological model Soil and Water Assessment Tool (SWAT), using ensemble of four high resolution (~25 km) Regional Climate Model (RCM) simulations for the near future (2021-2050) based on the SRES A1B scenario. Finally, to quantify the hydrological alterations of different flow characteristics, the Indicators of Hydrological Alteration (IHA) program based on the Range of Variability Approach (RVA) is used. This approach enables the assessment of ecologically sensitive streamflow parameters for the pre- and post-impact periods in the regions where availability of long-term ecological data is a limiting factor. Results indicate that flow variability has been significantly reduced due to dam construction with high flows being absorbed and pre-monsoon low flows being enhanced by the reservoir. Climate change alone

29 may reduce high peak flows while a combination of dam and climate change may significantly
30 reduce variability by affecting both high and low flows, thereby further disrupting the
31 functioning of riverine ecosystems. We find that, in the Kangsabati River basin, influence of dam
32 is greater than that of the climate change, thereby emphasizing the significance of direct human
33 intervention.

34

35 *Keywords: Anthropogenic impact, climate change, flow alteration, IHA, RCM, SWAT*

37 **1 Introduction**

38 Flow regime alteration of important seasonal flow components, such as high flows and low flows,
39 by anthropogenic activities, especially large dams, has generated immense scientific interest with
40 regards to implications for riverine ecosystems, biodiversity conservation and invasion by non-
41 native species (Bunn and Arthington, 2002; Lytle and Poff, 2004; Meijer et al. 2014).
42 Degradation of ecological health is now associated with the downstream section of dams (Poff
43 and Zimmermann 2010; Suen 2011). Carlisle et al. (2010) reported that across regions and
44 anthropogenic conditions, biological impairment is directly related to the magnitude of
45 streamflow reduction. Moreover, regulation of river flow and alteration of flood and drought
46 timing is expected to favour species that spawn during certain times (Freeman et al. 2001).

47 Along with direct anthropogenic impacts, human-induced climate change is also expected to
48 affect the hydrologic cycles and thereby alter natural flow characteristics. Increasing
49 temperatures will directly increase evaporation and alter plant transpiration rates, thereby
50 reducing runoff (Bates et al. 2008). Doll and Zhang (2010) have shown that by mid-21st century,
51 climate change effect on flow regimes may be greater than that caused by dams and water
52 withdrawals. Global analysis of potential changes in runoff regimes shows that by the year 2050,
53 most regions will experience significant changes in hydrological regime (Arnell and Gosling
54 2013). Changes brought about by climate change will interact with existing anthropogenic
55 factors and thus cause additional stress to riverine ecosystems (Fung et al. 2013; Ravazzani et al.
56 2015).

57 Much needed interaction between scientists from hydrological, ecological and geomorphological
58 foci over the past 20 years has increased our understanding of riverine dynamics, which is an
59 essential prerequisite for gauging future implications of human actions. Such studies typically
60 require long-term monitoring and assessment of baseline conditions to benchmark the effect of
61 changes (Wagener et al. 2010). However, increasingly, resources for developing such
62 quantitative understanding and data are declining (Mishra and Coulibaly 2009). Shifting baseline
63 conditions due to human intervention has added to the existing issue of insufficient ecological
64 information (Wagener et al. 2010). Historically, insufficient resources for regular survey and
65 assessment of ecological conditions of riverine systems have been a significant limitation for
66 carrying out change detection studies in developing countries.

67

68 In this study of the Kangsabati River basin, we address two important research gaps related to
69 natural flow regime alteration; (i) effect of anthropogenic activity (damming) and future climate
70 change for a mesoscale river basin with a strong monsoonal influence on hydrology and (ii)
71 usage of sparse and scattered ecological data to derive inferences regarding potential impacts of
72 damming and climate change on riverine ecosystem. We quantify observed alterations in the
73 flow regime due to damming and then model the ramifications of climate change using the
74 conventional ‘top-down’ hydrological modelling approach forced by (Regional Climate Model)
75 RCM simulations for the mid-21st century period. The study approach makes three novel
76 contributions to the existing body of knowledge.

- 77 • Few gauging stations in the developing world have long-term and accessible observed
78 discharge data which can be used for determining impact of a dam constructed 50 years
79 ago. This study is valuable because it extends our understanding of observed changes in
80 river flow regime in a developing country context.
- 81 • A methodological innovation in the modelling approach is that we examine the potential
82 impact of climate change alone by isolating the climate change signal. We also compare
83 potential future climate change impacts with combined impact of dam and climate change.
- 84 • For a mesoscale river basin, GCM outputs are not useful because they do not provide the
85 necessary spatial variability, which RCM simulations provide. The four RCM simulations
86 used here represent the most comprehensive set of high resolution future climate
87 simulations available for this region, which make them useful for assessing potential
88 scenarios of future climate change impact on the river flow regime.

89

90 **1.1 Description of the study area**

91 The Kangsabati River (basin area: 5,796 km²) originates in the Chotanagpur plateau of central
92 India, flows in a southeasterly direction to merge with the Ganges River in India, as its last
93 contributing river (Figure 1). Upper reaches have hardpan sub-surface geology while the middle
94 reaches consist of transitional undulating terrain, which levels out into the alluvial plains of the
95 lower reaches. The geology of this lateritic region and the excessively drained topography cause
96 high monsoon runoff coupled with low flow conditions during the dry months. Therefore,
97 despite a high average annual rainfall (western part, 1300 mm and eastern part, 1600 mm), the

98 basin has been traditionally considered drought prone due to low water holding capacity of the
99 lateritic soil, high summer temperatures and high evapotranspiration rates (Mishra and Desai,
100 2005; Saxena, 2012).

101 The Kangsabati reservoir is located at the confluence of the Kangsabati River and a major
102 tributary, Kumari. A dam constructed in 1965 on the Kangsabati River was followed by a second
103 connected dam over Kumari River in 1973. In the intermediate period, partial regulation of the
104 total flow took place. Since 1974 inflow to the Kangsabati reservoir comprises of the combined
105 streamflow of Kangsabati and Kumari sub-basins. The diverted water is primarily used for
106 irrigation in the reservoir command, the area of which is approximately 5,568 km². The dam also
107 provides flood water storage to mitigate the flooding problems in the lower reaches. High water
108 demand in the command area has also led to over-exploitation of groundwater resources and
109 consequently affected the river flow.

110

111 Figure 1

112

113 This river sustains the natural ecosystems which provides locals with their staple food; fish. It
114 also has the most diverse macrophytic riverine vegetation in the region with up to 80 species
115 found across the pre-monsoon, monsoon and post-monsoon seasons (Pradhan et al. 2005). Most
116 siluroid fishes in the region are commercially important and the lower reaches of the Kangsabati
117 River possess the greatest variety of fishes in the region. However, these fishes are highly
118 vulnerable to environmental degradation, particularly habitat destruction (Giri et al. 2008). The
119 studies performed in this region, being sporadic and short term, do not allow for a coherent long-
120 term ecosystem analysis of river discharge and ecological health.

121 Figure 2 presents the observed discharge at Mohanpur gauging station for the period 1950-2010,
122 where the 1950-1965 represents the natural flow regime, 1965-1973 represents partial effect of
123 dam, while dam altered flow regime prevails from 1974-2010. Barring the 1978 floods, the dam
124 has effectively kept peak flood levels below the 4000 m³/s mark. The dampening effect of the
125 dam is also clearly visible with larger bases of the flood peaks after 1985. Beyond existing
126 anthropogenic interventions, impending climate change is expected to alter the hydrological
127 characteristics of the region by reducing the frequency of extreme precipitation events and
128 lengthening dry spells (Mittal et al. 2013).

129

130 Figure 2

131

132 **1.2 Study design**

133 Assessment of ecologically important natural flow regime characteristics necessitates long-term
134 data, especially for the period prior to the onset of an impact event or change. The gauging
135 station at Mohanpur, about 80 km downstream of the reservoir has pre-dam discharge data for
136 the period 1950-1965, which may be considered enough for a bias-free and appropriate
137 assessment (Kennard et al. 2010). After the intermediate period of 9 years (1966-1973), where
138 the influence of damming is partial and therefore difficult to understand in terms of impact, a
139 total of 37 years of post-dam discharge information is available (1974-2010). This constitutes the
140 observed data and forms the basis for the pre- and post-dam analysis at Mohanpur. Variability in
141 regulated rivers is highly influenced by water use, while climatic forcing at different time scales
142 also brings about hydrological changes. Therefore, it is crucial to separate flow regime changes
143 caused by climate change from dam effects, so that a better knowledge of ecosystem impacts and
144 potential restoration may be developed (Zolezzi et al. 2009). Based on this understanding,
145 analysis of impact of dam and climate change on streamflow has been carried out in three parts;
146 (i) effect of dam (ii) impact of future climate change (climate change signal) and (iii) impact of
147 both dam and climate change in the future. Hydrologic alteration of biologically relevant flow
148 regimes expected to be caused by dam construction and climate change are assessed using
149 Indicators of Hydrologic Alteration (IHA) (The Nature Conservancy 2009).

150

151 **2. Methods and Data**

152 **2.1 SWAT hydrologic model**

153 SWAT 2009 (Neitsch et al. 2009) is used to simulate river discharges for observed and future
154 period. SWAT typically operates on a daily time step and accounts for spatial heterogeneities of
155 soil, land cover and elevation, by subdividing basin into multiple hydrological response units
156 (HRUs). The rainfall-runoff model simulates the discharge from each sub basin and routes the
157 streamflow to the watershed outlet (Neitsch et al. 2009). Preprocessing and model setup were
158 performed using the Arc-SWAT extension for ArcGIS 9.3. The Sequential Uncertainty Fitting
159 algorithm (SUFI-2) (Abbaspour et al. 2007) is used to calibrate SWAT and quantifies uncertainty

160 using P factor and R factor statistics. The P factor, which varies from 0 to 1, represents the
161 fraction of observed discharge which falls within the 95PPU band, while the R factor is derived
162 by taking the ratio of the average width of the 95PPU and the standard deviation of the observed
163 discharge. While a value of less than 1 is considered desirable for R factor, the ideal value for P
164 factor is 1 (100% values within the band) (Vaghefi et al. 2013). 95PPU is 95 Percent Prediction
165 Uncertainty, calculated at the 2.5% and 97.5% levels of an output variable, disallowing 5% of
166 the bad simulations. Three evaluation criteria are used to assess model performance: Percent bias
167 (PBIAS), Nash-Sutcliffe efficiency (NSE) and coefficient of determination (R^2). PBIAS, NSE
168 and R^2 describe the goodness-of-fit between simulated and observed flow; and the model
169 simulation would be considered satisfactory when PBIAS values are $< 25\%$ and best when their
170 values approach one in case of NSE and R^2 (Moriasi et al. 2007).

171

172 **2.2 Assessment of hydrologic alteration**

173 IHA methodology based on Range of Variability Approach (RVA) is applied to assess the degree
174 of departure from natural flow regime that has already occurred due to dam construction and is
175 expected in the future due to climate change (Richter et al. 1997). RVA is the most widely used
176 approach for quantifying hydrologic alterations in order to set appropriate environmental flow
177 targets (Zolezzi et al. 2009). To analyse the degree of hydrologic alteration in ecologically
178 relevant statistics, a subset of indices is used, as there exists redundancy among the indices
179 representing different flow components (Olden and Poff 2003).

180 For RVA analysis, the pre-impact streamflow data is divided into three different categories;
181 values upto 33rd percentile (lower category), 34th to 67th percentile (middle category) and values
182 greater than 67th percentile (high category). A Hydrologic Alteration factor is calculated for each
183 of the three categories as: (observed frequency – expected frequency) / expected frequency. A
184 positive Hydrologic Alteration (HA) value indicates an increase in frequency of values in the
185 category while negative indicates a reduction. In the absence of specific ecological information,
186 the range between the 34th and the 67th percentile, i.e. the middle category is identified as the
187 targeted range of variability for the post-impact period.

188

189 **2.3 Observed input data**

190 SWAT model required input for topography, soil and land use/land cover which are compiled
191 from Global Land Cover Facility (GLCF) website, National Bureau of Soil Survey and Land Use
192 Planning (NBSS&LUP), unsupervised classification of digital remote sensing images of Landsat
193 5 Thematic Mapper (TM) for year 1990 (dated 07/11/1990 and 21/11/1990) and Landsat 7
194 Enhanced Thematic Mapper (ETM+) for year 2001 (dated 26/10/2001 and 02/11/2001)
195 respectively. Observed climate data including precipitation, maximum air temperature and
196 minimum air temperature from 1991 to 2010 for five weather stations (Figure 1) are gathered
197 from India Meteorological Department (IMD) and Agro-Meteorology Department, Government
198 of West Bengal. Observed discharge data from river gauging stations, Simulia, Tusuma,
199 Rangagora, Kharidwar and Mohanpur are collected from the Central Water Commission (CWC)
200 and Irrigation and Water Ways Department (IWWD), Government of West Bengal. The
201 Kangsabati reservoir is included with reservoir operational information starting from 1974, when
202 the second phase of Kangsabati dam completed. Reservoir management information includes
203 measured monthly outflow to calculate reservoir outflow, reservoir surface area when reservoir
204 is filled to emergency (12498 ha) and principal spillway (11101 ha), volume of water needed to
205 fill the reservoir to the emergency (123500 m^3) and principal spillway (98186 m^3).

206

207 **2.4. Future climate data**

208 Daily precipitation, maximum and minimum temperature from four RCM simulations and their
209 ensemble mean are used to drive calibrated SWAT. The historical (control) simulations for the
210 period 1970-1999 and A1B SRES emission scenario based future climate simulations for the
211 period 2021-2050 from four RCM simulations, REMO-ECHAM5, REMO-HadCM3, HadRM3-
212 ECHAM5 and HadRM3-HadCM3; are obtained by the forcing from two CMIP3 GCMs namely
213 ECHAM5-MPIOM and HadCM3 and two RCMs; REMO and HadRM3. The performance of
214 these RCMs for the Kangsabati basin has been validated by comparing 20 year model
215 simulations for the period 1989–2008, driven by lateral boundary forcings from ERAInterim
216 reanalysis data (Simmons et al. 2007), with the observational datasets; Climate Research Unit
217 (CRU) for temperature and Asian Precipitation Highly Resolved Observational Data
218 (APHRODITE) for precipitation. Both the RCMs have demonstrated an adequate ability to
219 capture the seasonal characteristics and interannual variability (IAV) of temperature and
220 precipitation (Mittal et al. 2013). The ensemble mean of four RCM simulations are used to

221 simulate future streamflow, due to which the use of bias correction is considered unnecessary
222 (Maurer and Pierce 2014). The use of ensemble reduces the uncertainties in climate projection
223 and provides more quantitative information for subsequent hydrologic impacts research (Jung et
224 al. 2012).

225

226 **3. Results and Discussion**

227 **3.1 SWAT model parameter sensitivity analysis**

228 SWAT model was calibrated for the Kangsabati river basin using monthly observed streamflow
229 at the five gauging station, during for the period 1991 to 2000. Due to the unavailability of
230 observed weather data for the pre-dam period from 1950 to 1965, SWAT calibration was carried
231 out using the post-dam period data. Initially, wide but meaningful ranges are assigned to
232 sensitive parameters and with further simulations final ranges of model parameters were
233 determined. The parameters with highest sensitivity are used to calibrate and validate the model.
234 Table 1 shows the sensitive parameters included in the final calibration, their initial ranges,
235 initial and final values and their t and p values. Eleven parameters representing the surface runoff,
236 groundwater and soil properties are found to be sensitive in the estimation of streamflow. t-
237 statistics provides a measure of sensitivity (larger in absolute values are more sensitive) and p-
238 values determined the significance of the sensitivity with a values close to zero having more
239 significance. Having high t-statistics and low p-value; Curve Number (CN2), alpha baseflow
240 (ALPHA BF) and groundwater delay (GW DELAY) parameters are found to be the most
241 sensitive to streamflow.

242

243 Table 1

244

245 **3.2 SWAT model calibration, validation and uncertainty analysis**

246 The statistical comparison between observed and SWAT simulated streamflow at different
247 gauging stations during calibration period from 1991 to 2000 shows PBIAS values ranging from
248 -12.4 to 7.9%, higher values of R^2 (ranging from 0.66–0.87) and NSE (ranging from 0.63–0.74)
249 for all the gauging stations (Table 2). This suggests that model simulation can be judged as
250 satisfactory as PBIAS values range between the $\pm 25\%$ limits, R^2 is greater than 0.6 and NSE is
251 greater than 0.5 (Moriassi et al. 2007), although NSE values for Simulia and Kharidwar are below

252 0.65, considered to be an acceptable value (Ritter et al. 2013). The P factor indicates that for all
253 stations, more than 72% of the data are bracketed in the prediction uncertainty of the model,
254 whereas the R factors are mostly around 1 except Mohanpur gauging station where the P factor is
255 40% and R factor is 0.59.

256 For validation for the period 2001 to 2010, PBIAS, R^2 and NSE validation values ranges from -
257 4.8 to 11.8%, 0.66 to 0.85 and 0.53 to 0.76 respectively, indicating a good relationship between
258 observed and simulated streamflow values except for Simulia, Kharidwar and Mohanpur station
259 with low NSE values of 0.53, 0.64 and 0.49 which are unsatisfactory according to Ritter et al.
260 (2013). The P factor indicates that for all stations, more than 62% of the data are bracketed in the
261 prediction uncertainty of the model, whereas the R factors are mostly around or below 1 except
262 Tusuma gauging station where the R factor is 0.65. In general, in the downstream of Kangsabati
263 dam, the model prediction has larger uncertainties. Poor calibration and validation results in case
264 of managed streamflow have also been observed before (Faramarzi et al. 2010; Vaghefi et al.
265 2013).

266

267 Table 2

268

269 **3.3 SWAT model simulation**

270 The calibrated model is used to simulate streamflow for two time periods, 1970-1999 (control)
271 and 2021-2050 (future), based on ensemble mean of four RCM simulations for the SRES A1B
272 scenario. To analyse the impact of climate change and the combined effect of dam and climate
273 change, two separate simulations are carried out.

274 *Simulation 1* (impact of climate change) - SWAT model streamflow simulations for these control
275 period simulations without the inclusion of the Kangsabati dam represent the natural flow regime
276 of the basin. Comparison of this flow regime with SWAT simulated flow regime for the future
277 period (2021-2050) is used to isolate the impact of climate change.

278 *Simulation 2* (impact of dam and climate change) - SWAT model is run for the future period
279 (2021-2050) based on the RCM simulations and their ensemble. Kangsabati dam is included in
280 this simulation to analyse the streamflow conditions due to both, dam and climate change.
281 Comparison of future period simulations with observed streamflow for pre-dam period (natural

282 flow regime - 1950-1965) is used to assess the combined impact of dam and climate change on
283 the natural flow regime.

284

285 **3.4 Impact of dam on flow regime**

286 The primary function of the Kangsabati dam is to divert water for irrigation and to mitigate the
287 impacts of monsoon floods. The IHA based analysis is described from the perspective of pre-
288 monsoon, monsoon and post-monsoon periods. Seasonal variations in flow of the river after dam
289 construction are much relevant to the physiological and life cycle stages of various freshwater
290 fishes. In this case, *Bagarius bagarius*, the largest freshwater migratory siluroid fish (catfish),
291 which is abundantly present in rivers flowing through West Bengal, is found to be absent in the
292 downstream section of the Kangsabati dam ((Hamilton, 1822, Mishra and Coulibaly 2009). It is
293 categorized threatened by the International Union for Conservation of Nature (IUCN 2013),
294 primarily due to its decline as a result of dam construction which prevents their upstream
295 migration for spawning (Lakra et al. 2011). The effect of Kangsabati dam on the observed flow
296 regime is depicted in Figure 3, through monthly average flows, monthly low flows and Flow
297 Duration Curves (FDCs) for the representative months of April (pre-monsoon), July (monsoon)
298 and November (post-monsoon). During pre-monsoon, the middle value for monthly average and
299 monthly low flows is higher in the post dam period, largely due to periodic dam releases during
300 the otherwise dry period characterized by natural minimum flows. Post-impact period is also
301 characterized by greater flow variability, with more frequent high flow events. The
302 corresponding FDC clearly corroborates this assessment, by depicting persistent higher flow
303 rates for more than 80% of the time period as well as significantly lower flow rates for the
304 remaining 20% of the time. Whereas in the monsoon season, the dam dampens the monthly
305 average and low flows by absorbing high flow pulses and maintaining a more consistent flow
306 rate. The FDC clearly demonstrates the overall effect, where the difference between the area
307 under the curves for the pre-dam and post-dam periods corresponds to the amount of water
308 diverted for irrigation purpose. Irrigation requirements for the Rabi (winter season) crop further
309 reduce the discharge downstream of the dam during the post-monsoon month of November. A
310 fraction of the high flows is diverted for this purpose, thereby reducing the monthly average
311 flows.

312

313 Figure 3

314

315 **3.5 Effect of climate change on flow regime**

316 Figure 4 (a) show the effect of climate change on simulated flow regime through FDC, EFCs and
317 hydrologic alteration graphs based on the output of SWAT “Simulation 1- impact of climate
318 change” for the time period 2021-2050. The comparative analysis of FDCs in Figure 4 reveals
319 the effect of the climate change vis-à-vis the combined effect of dam and climate change. In this
320 case, as the FDC demonstrates, climate change reduces flows, but the area under the curve is
321 affected to a lesser degree than for the impact of dam alone (Section 3.4).

322 Figure 4 (b) demonstrates the deviation factor of coefficients of dispersion (CD), which
323 represents the change in flow variability as represented by EFCs during the mid-21st century
324 compared to control period. Climate change causes deviation in both extreme low flow and high
325 flow components. Deviation for high flow peak and frequency is higher (>0.6), but the deviation
326 for extreme low flow peak, duration, timing and frequency is lower (<0.6). A slight change in the
327 timing of high flow pulses affects the benthic siluroid fishes which are very good indicators of
328 habitat degradation (Wootton et al. 1996). The change in flow timings affects their life cycle by
329 disrupting various stages such as spawning, egg hatching, rearing, movement onto the
330 floodplains for feeding and reproduction or migration upstream and downstream (Poff et al.
331 1997).

332 Figure 4 (c) shows the extent of Hydrologic Alteration (HA) in monthly flows during the mid-
333 21st century. A high HA (>0.5) for the high and low category is projected for the monsoon
334 months (JJAS) while rest of the months show less hydrologic alteration in both the categories.
335 Months of May and September shows high alteration in both the middle and high category,
336 whereas less alteration is observed in all three categories during post-monsoon (ON) and winter
337 months (DJF). This reduction may significantly affect the connectivity with the flood plains by
338 potentially reducing the magnitude and areal spread of floods. Such drastic changes will affect
339 the yolk-sac-larva of threatened species of siluroid fish *Mystus gulio*, which develops in
340 floodplain freshwater (IUCN, 2013; Termvidchakorn and Hortle, 2013). Along with decrease in
341 the number of high flow events, reduction of high flow duration and changes in their timing will
342 add additional physiological stress to the fish species (Sharma and Shrestha, 2001). As of now,
343 Mishra and Coulibaly (2009) reported a decline of 27.8% in *Mystus gulio* catch across

344 southwestern Bengal. Observed reduction may be due to a combination of stressors such as
345 overfishing, flow alteration and habitat loss, but in the absence of biological information and
346 temporal monitoring of stressors, the influence of individual factors cannot be determined
347 (Sarkar and Bain, 2007).

348

349 Figure 4

350

351 **3.6 Effect of dam and climate change on flow regime**

352 The output of SWAT “simulation 2 – dam and climate change” is used to analyse the combined
353 impact of dam and climate change on hydrologic indicators in the Kangsabati basin during the
354 mid-21st century (2021-2050) compared to pre-dam period (1950-1965). The combined effect of
355 dam and climatic changes, depicted in Figures 4 indicates significant reduction in the magnitude,
356 frequency and duration of extreme high and medium flow rates in the simulated flow, whereas, a
357 small increase in low flows is observed in comparison with sole effect of climate change.

358 The FDC shows how extreme high flows above 2000 m³/s are eliminated in this scenario. The
359 natural flow regime shows a consistent temporal distribution with ~ 75% flows lying in the range
360 from 200 m³/s to 10 m³/s (Figure 4 (a)). However, in the altered future condition, this percentage
361 reduces significantly to ~10%. Similarly, the combined effect of dam and climate change shows
362 significantly greater alteration in EFCs compared to only climate change.

363 Figure 4 (b) represents the deviation factor in CD of EFCs during future (2021-2050) period in
364 comparison with natural flow (1950-1965). In case of combined impact of dam and climate
365 change, deviation of >0.5 is observed for CD of all extreme low flow and high flow
366 characteristics, as compared to the individual impact of climate change. There is also a moderate
367 increase in deviation in high flow frequency and duration due to the combined effect of dam and
368 climate change. Significant changes in timing, frequency and duration of extreme low and high
369 flows implies that the life cycle of many aquatic species may get disrupted during various stages
370 such as spawning, egg hatching, rearing and their movement onto the floodplains for feeding and
371 reproduction (Suren and Riis 2010). Benthic siluroid fishes found commonly in the Kangsabati
372 River, which have declined since the 1960’s (Mishra and Coulibaly 2009), are highly sensitive to
373 reduction in high flows which cause habitat degradation through channel bed sedimentation
374 (Lisle 1989).

375 Figure 4 (c) shows the extent of HA in monthly flows due to the combined effect of dam and
376 climate change. Unlike previous two scenarios, individual impact of dam and climate change;
377 high hydrologic alteration in either of high, middle or low alteration category is distributed
378 throughout the year except for the month of October, where HA is comparatively less. Positive
379 HA (>0.5) in the high category is projected for June, August, September and November months
380 while negative HA (> -0.5) is observed in January, February, March, April, July and December.
381 Low flows in the month of January, February, March and April are projected to increase due to
382 the increase in HA in the low category in these months whereas conversely shows reduction
383 during May, June and November months.

384 Previous global analysis has indicated that the impact of climate change on flow regime is larger
385 than the effect of dams and water withdrawals (Doll and Zhang 2010). However, this may be on
386 account of two factors; the likely underestimation of dam impacts (Doll and Zhang 2010) and the
387 high degree of spatial variation and basin specific impacts. An important factor which needs
388 consideration is that the Kangsabati reservoir storage represents about one-third of the total
389 annual discharge and is, therefore, a major factor in altering the flow regime of this basin. Future
390 climate change will, therefore, put additional stress leading to greater risk of ecological change in
391 a riverine ecosystem already affected by anthropogenic interference.

392

393

394 **4. Conclusions**

395 This study provides a detailed basin scale assessment of ecologically relevant flow alterations in
396 a monsoon dominated, drought prone river basin. IHA and EFC parameters describing changes
397 in long-term monthly average, timing, duration and frequency of extreme flow conditions show a
398 significant change from the natural flow regime during the observed period. Dampening effect of
399 dams on hydrological variability and the extreme seasonality of river flows is highly pronounced.
400 Significant overall flow reduction by the dam for provision of irrigation and domestic water
401 demands will be exacerbated by climate change. The combined effect of dam and climate change
402 is found to be significantly greater than the individual impact of dam or climate change.
403 However, lack of sufficient long-term ecological data is a limitation in the assessment of habitat
404 changes in the Kangsabati basin. We find that the ecologically sensitive IHA parameters and the
405 associated inferences that may be drawn regarding the impacts on aquatic species are useful in

406 cases where availability of long-term ecological data is a drawback. There is an urgent need to
407 correlate real time ecological, bio-geochemical and morphological characteristics with observed
408 hydrological changes to better assess the vulnerability of aquatic ecosystems to future changes. A
409 better understanding of ecosystem impacts will be useful to inform the method of river
410 restoration and ecosystem management programmes in the future.

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431 **References**

- 432 Abbaspour KC, Yang J, Maximov I, Siber R, Bogner K, Mieleitner J, Zobrist J, Srinivasan R
433 (2007) Modelling hydrology and water quality in the pre-alpine/alpine Thur watershed using
434 SWAT. *J Hydrol* 333: 413-430
- 435 Arnell NW, Gosling SN (2013) The impacts of climate change on river flow regimes at the
436 global scale. *J Hydrol* 486: 351–364
- 437 Bates B, Kundzewicz ZW, Wu S, Palutikof J (2008) Climate change and water.
438 Intergovernmental Panel on Climate Change (IPCC)
- 439 Bunn SE, Arthington AH (2002) Basic principles and ecological consequences of altered flow
440 regimes for aquatic biodiversity. *Environ Manage* 30: 492-507
- 441 Carlisle DM, Wolock DM, Meador MR (2010) Alteration of streamflow magnitudes and
442 potential ecological consequences: a multiregional assessment. *Front Ecol Environ* 9: 264-270
- 443 Döll P, Zhang J (2010) Impact of climate change on freshwater ecosystems: a global-scale
444 analysis of ecologically relevant river flow alterations. *Hydrol Earth Syst Sci Discuss* 7: 1305-
445 1342
- 446 Faramarzi M, Yang H, Mousavi J, Schulin R, Binder CR, Abbaspour KC (2010) Analysis of
447 intra-country virtual water trade strategy to alleviate water scarcity in Iran. *Hydrol Earth Syst Sci*
448 *Discuss*, 7(2)
- 449 Freeman MC, Bowen ZH, Bovee KD, Irwin ER (2001) Flow and habitat effects on juvenile fish
450 abundance in natural and altered flow regimes. *Ecol Appl* 11: 179-190
- 451 Fung F, Watts G, Lopez A, Orr HG, New M, Extence C (2013). Using large climate ensembles
452 to plan for the hydrological impact of climate change in the freshwater environment. *Wat.*
453 *Resour. Mgmt.*, 27(4), 1063-1084.
- 454 Giri A, Pradhan P, Chakraborty SK (2008) Studies on hydrobiological status of Kansai and
455 Dwarkeswar rivers in West Bengal, India. *J Inland Fish Soc* 40: 59-64
- 456 Hamilton F (1822) Account of the fishes found in the river Ganges and its branches. Edinburg,
457 London. 1-405
- 458 IUCN (2013) Red List of Threatened Species. Version 2013.2. www.iucnredlist.org
- 459 Jung IW, Bae DH, Lee BJ (2012) Possible change in Korean streamflow seasonality based on
460 multi-model climate projections. *Hydrol Process*. doi: 10.1002/hyp.9215
- 461 Kennard MJ, Pusey BJ, Olden JD, MacKay SJ, Stein JL, Marsh N (2010) Classification of
462 natural flow regimes in Australia to support environmental flow management. *Freshw Biol* 55:
463 171-193
- 464 Lakra WS, Sarkar UK, Dubey VK, Sani R, Pandey A (2011) River inter linking in India: status,
465 issues, prospects and implications on aquatic ecosystems and freshwater fish diversity. *Rev Fish*
466 *Biol Fisher* 21: 463-479

467 Lisle TE (1989) Sediment transport and resulting deposition in spawning gravels, north coastal
468 California. *Water Resour Res* 25: 1303 – 1319

469 Lytle DA, Poff NL (2004) Adaptation to natural flow regimes. *Trends Ecol Evol* 19: 94-100

470 Maurer EP, Pierce DW (2013) Bias correction can modify climate model-simulated precipitation
471 changes without adverse affect on the ensemble mean. *Hydrol Earth Syst Sc* 18: 915-925

472 Meijer KS, Van der Krogt WNM, Van Beek E (2012). A new approach to incorporating
473 environmental flow requirements in water allocation modeling. *Wat. Resour. Mgmt.*, 26(5),
474 1271-1286.

475 Mishra AK, Coulibaly P (2009) Developments in hydrometric network design: A review. *Rev*
476 *Geophys* 47(2)

477 Mishra AK, Desai VR (2005) Drought forecasting using stochastic models. *Stoch Environ Res*
478 *Risk Assess* 19: 326-339

479 Mittal N, Mishra A, Singh R (2013) Combining climatological and participatory approaches for
480 assessing changes in extreme climatic indices at regional scale. *Clim Change* 1-13

481 Moriasi DN, Arnold JG, Van Liew MW, Bingner RL, Harmel RD, Veith TL (2007) Model
482 evaluation guidelines for systematic quantification of accuracy in watershed simulations. *T*
483 *ASAE* 50: 885-900

484 Neitsch SL, Arnold JG, Kiniry JR, Williams JR (2009) Soil and Water Assessment Tool:
485 Theoretical Documentation, Version 2009. URL: <http://twri.tamu.edu/reports/2011/tr406.pdf>

486 Olden JD, Poff NL (2003) Redundancy and the choice of hydrologic indices for characterizing
487 streamflow regimes. *River Res Appl* 19: 101-121

488 Poff NL, Zimmerman JK (2010) Ecological responses to altered flow regimes: a literature review
489 to inform the science and management of environmental flows. *Freshw Biol* 55: 194-205

490 Poff NL, Allan JD, Bain MB, Karr JR, Prestegard KL, Richter BD, Sparks RE, Stromberg JC
491 (1997) The natural flow regime. *BioSci* 47: 769-784

492 Pradhan P, Mishra SS, Chakraborty SK, Bhakat RK (2005) Diversity of freshwater macrophytic
493 vegetation of six rivers of south West Bengal. *Trop Ecol* 46: 193-202

494 Ravazzani G, Barbero S, Salandin A, Senatore A, Mancini M (2014). An integrated hydrological
495 model for assessing climate change impacts on water resources of the upper Po river basin. *Wat.*
496 *Resour. Mgmt.*, 29(4), 1193-1215.

497 Richter BD, Baumgartner JV, Wigington R, Braun DP (1997) How much water does a river need?
498 *Freshw Biol* 37: 231-249

499 Ritter A, Muñoz-Carpena R (2013) Performance evaluation of hydrological models: Statistical
500 significance for reducing subjectivity in goodness-of-fit assessments. *J Hydrol* 480: 33-45

501 Sarkar UK, Bain MB (2007) Priority habitats for the conservation of large river fish in the
502 Ganges river basin. *Aquatic Conserv Mar Freshw Ecosyst* 17: 349-359

503 Saxena RP (2012) Impacts of Kangsabati Project, India. In *Impacts of Large Dams: A Global*
504 *Assessment (277-298)*. Springer Berlin Heidelberg

505 Sharma CM, Shrestha J (2001) Fish diversity and fishery resources of the Tinau River, Western
506 Nepal. *Environment and Agriculture: Biodiversity, Agriculture and Pollution in South Asia*, p78-
507 83, Ecological Society (ECOS), Kathmandu, Nepal.

508 Simmons AS, Uppala DD, Kobayashi S (2007) ERA-interim: new ECMWF reanalysis products
509 from 1989 onwards. *ECMWF Newsl* 110:29–35

510 Suen, J. P. (2011). Determining the ecological flow regime for existing reservoir operation. *Wat.*
511 *Resour. Mgmt.*, 25(3), 817-835.

512 Suren AM, Riis T (2010) The effects of plant growth on stream invertebrate communities during
513 low flow: a conceptual model. *J N Am Benth Soc* 29: 711-724

514 Termvidchakorn A, Hortle KG (2013) A guide to larvae and juveniles of some common fish
515 species from the Mekong River Basin. MRC Technical Paper No. 38. Mekong River
516 Commission, Phnom Penh. 234pp. ISSN: 1683-1489

517 The Nature Conservancy (2009) Indicators of hydrologic alteration, version 7.1. User’s manual.
518 URL:<http://www.conservationgateway.org/ConservationPractices/Freshwater/EnvironmentalFlows/MethodsandTools/IndicatorsofHydrologicAlteration/Pages/IHA-Software-Download.aspx>
519

520 Vaghefi AS, Mousavi SJ, Abbaspour KC, Srinivasan R, Yang H (2013) Analyses of the impact
521 of climate change on water resources components, drought and wheat yield in semiarid regions:
522 Karkheh River Basin in Iran. *Hydrol Process*. doi: 10.1002/hyp.9747

523 Wagener T, Sivapalan M, Troch PA, McGlynn BL, Harman CJ, Gupta HV, Kumar P, Rao SC,
524 Basu NB, Wilson JS (2010) The future of hydrology: An evolving science for a changing world.
525 *Water Resour Res* 46(5): W05301

526 Wootton, J.T., M.S. Parker and M.E. Power. 1996. The effect of disturbance on river food webs.
527 *Science* 273: 1558-1561

528 Zolezzi G, Bellin A, Bruno MC, Maiolini B, Siviglia A (2009) Assessing hydrological
529 alterations at multiple temporal scales: Adige River, Italy. *Water Resour Res* 45(12)

Tables

Table 1 Sensitive SWAT parameters included in the final calibration, their initial ranges, final values and their t and p values

Parameter name ¹	Definition	t-statistics ²	p-value ²	Initial value	Range of values in SWAT-CUP	Final value
v__ALPHA_BF.gw	Base-flow alpha factor (days)	4.29	0.00	0.048	0.0-0.7	0.54
v__GW_DELAY.gw	Groundwater delay (days)	-7.65	0.00	31	0.0-250.0	80.75
v__GWQMN.gw	Threshold depth of water for return flow	-1.53	0.13	0	0.0-1.2	0.81
v__GW_REVAP.gw	Groundwater revap (water in the shallow aquifer returning to root zone) coefficient	1.43	0.15	0.02	0.0-0.2	0.10
v__ESCO.hru	Soil evaporation compensation factor	1.06	0.29	0.95	0.75-0.95	0.78
v__CH_N2.rte	Manning's N value for the main channels	-0.11	0.91	0.014	0.12-0.4	0.35
v__CH_K2.rte	Effective hydraulic conductivity in main channel	-1.42	0.16	0	0.0-74.0	33.37
Parameter name ¹	Definition	t-statistics ²	p-value ²	Initial value	Initial range of multiplier in SWAT-CUP	Final value of the multiplier
r__CN2.mgt	SCS runoff curve number for moisture condition II	11.31	0.00	75-98	-0.4-0.004	-0.18
r__SOL_AWC.sol	Available water capacity of first soil layer (mm/mm)	0.35	0.72	0.06	0.0-0.4	0.39
r__SOL_K (1).sol	Saturated hydraulic conductivity of first soil layer (mm/h)	1.42	0.16	500	0.0-1.6	0.22
r__SOL_BD (1).sol	Moist bulk density of first soil layer (mg/m ³)	0.98	0.33	1	0.0-0.7	0.61

¹The qualifier (v_) refers to the substitution of a parameter by a value from the given range, while (r__) refers to a relative change in the parameter where the current values is multiplied by 1 plus a factor in the given range.

²The t-statistics and p-values are results from 500 runs of SUFI2 simulations; the larger t-statistics and smaller p-value, shows more sensitive parameter.

Table 2 SWAT Model performance of five calibrated subbasins in the Kangsabati basin.

River discharge station	Calibration (1991-2000)					Validation (2001-2010)				
	<i>P</i> factor	<i>R</i> factor	PBIAS	NSE	R ²	<i>P</i> factor	<i>R</i> factor	PBIAS	NSE	R ²
Simulia	0.74	1.08	7.9	0.63	0.69	0.71	0.94	11.8	0.53	0.69
Tusuma	0.78	0.75	-6.1	0.72	0.86	0.76	0.65	5.3	0.76	0.79
Rangagora	0.68	1.10	-12.4	0.74	0.66	0.62	1.02	-4.8	0.67	0.66
Kharidwar	0.72	0.84	5.8	0.64	0.75	0.79	0.90	8.3	0.64	0.75
Mohanpur	0.40	0.59	5.4	0.68	0.87	0.63	0.72	6.2	0.49	0.85

Figures

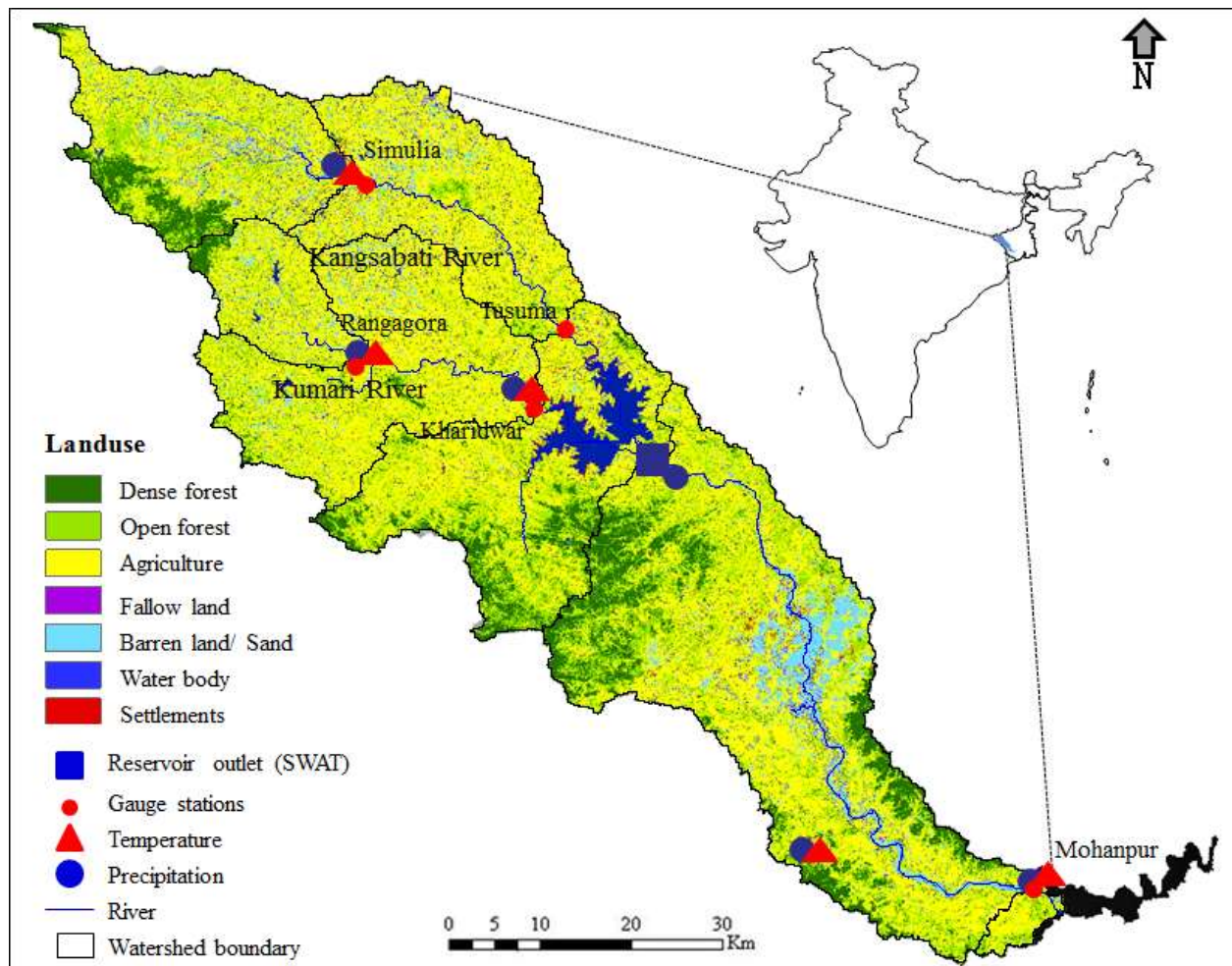


Figure 1

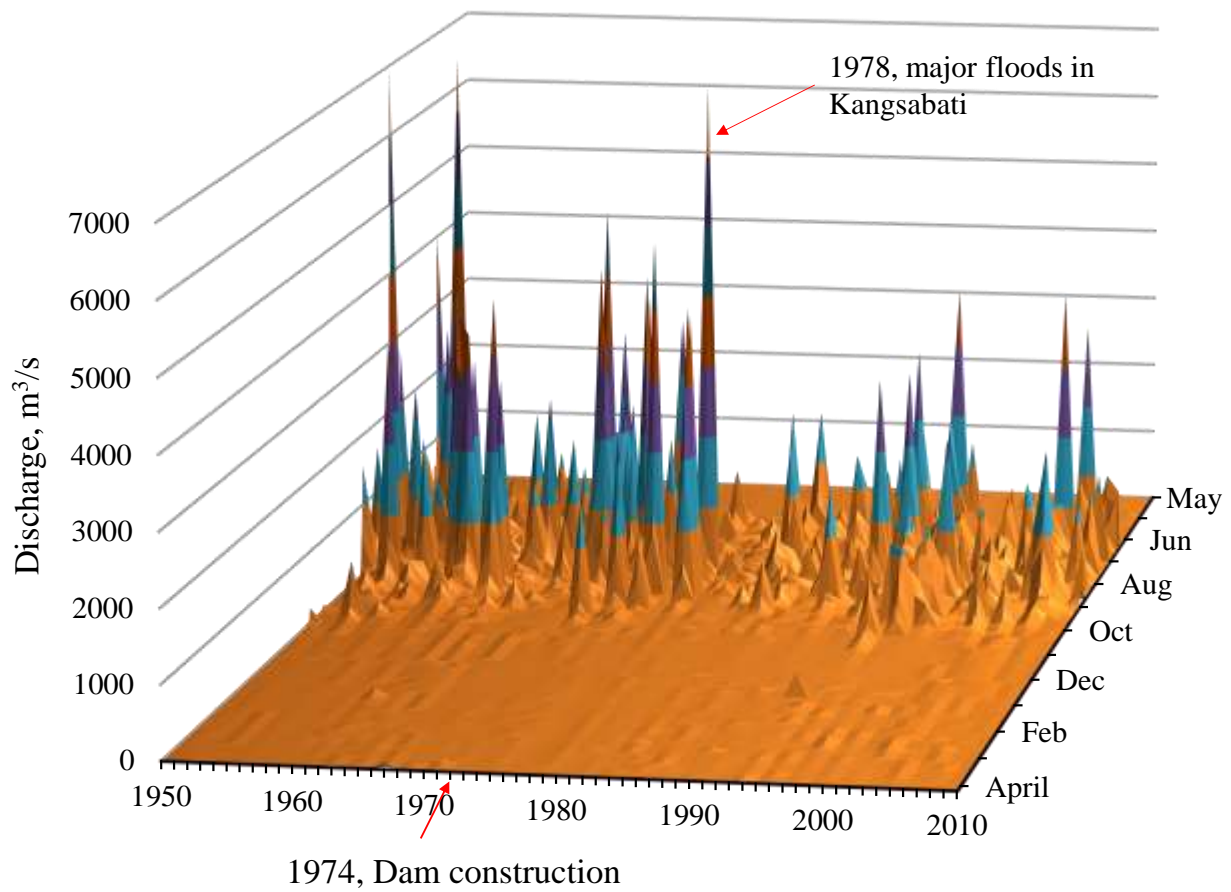


Figure 2

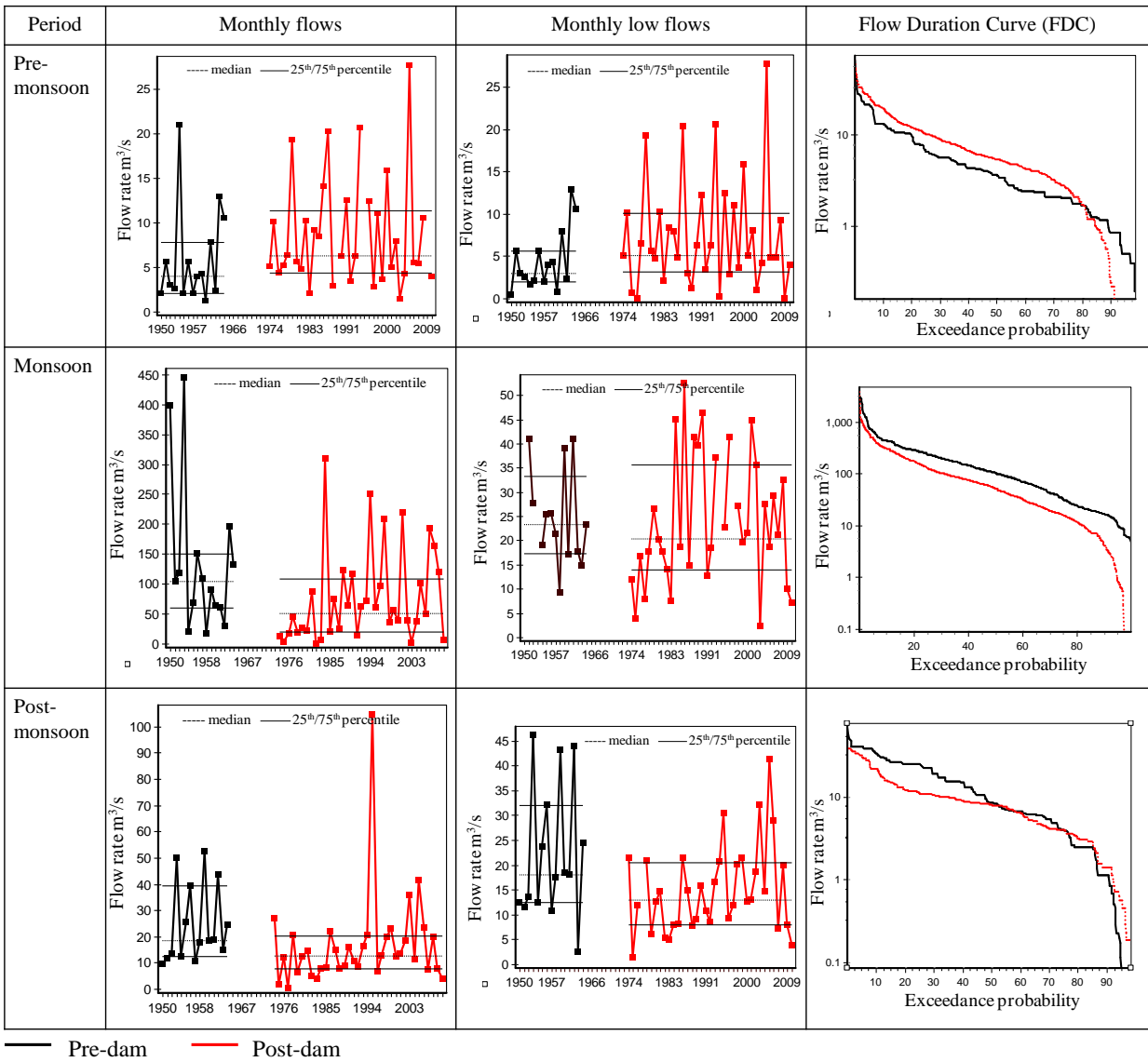


Figure 3

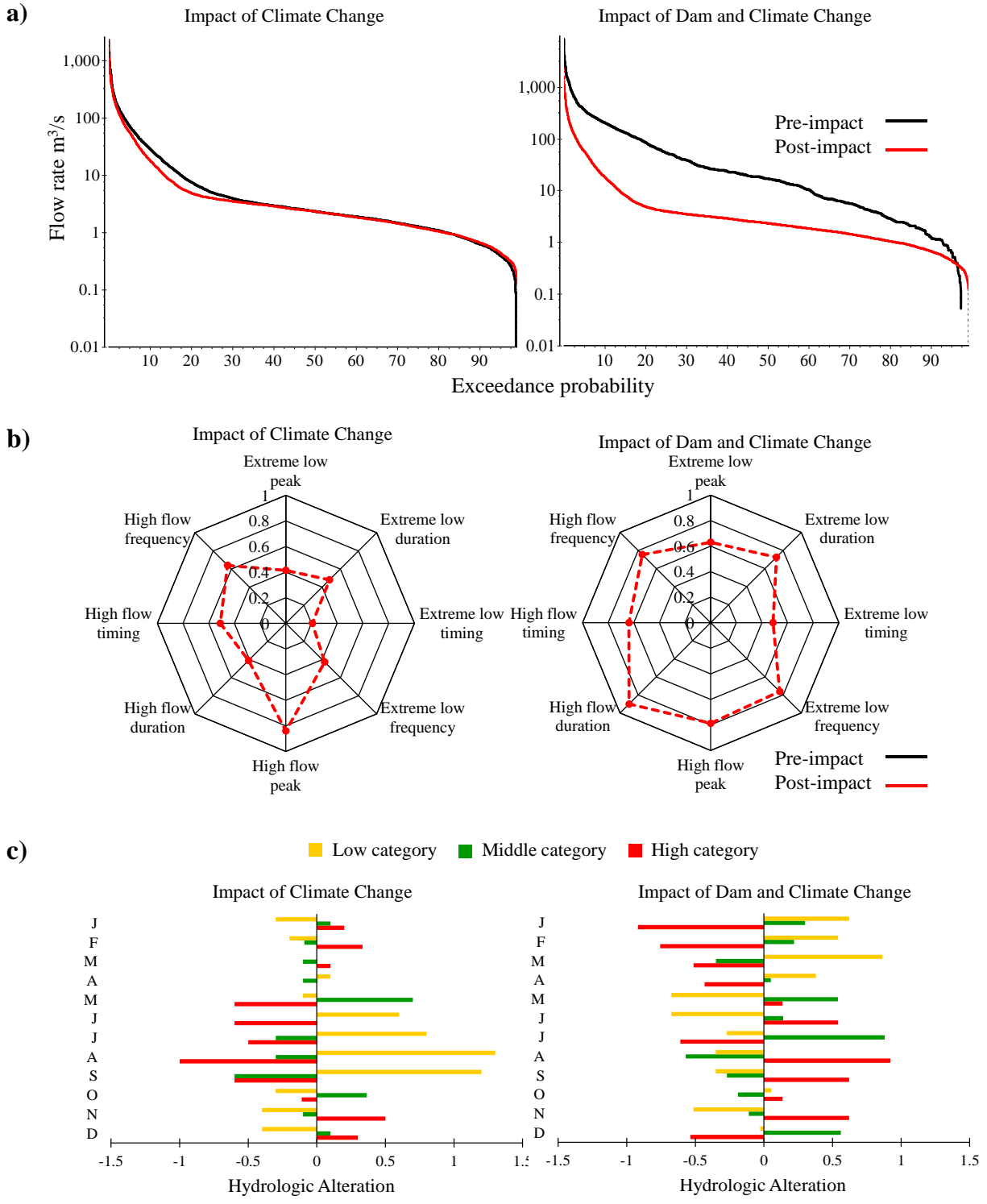


Figure 4